whiteion Bulletin, Vol. 11, pp. 191-195 Pages Press Ltd. 1980. Printed in Great Britain. 0025-326X/80/0701-0191\$02.00/0

## DT, DDE, and PCB Residues in Fish, trustaceans and Sediments from the lastern Mediterranean Coast of Turkey

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interest commercial bony fishes: grey mullet, red mullet, red mullet, and mullet and gold bandgoat fish, as well as in shrimps, and sediments obtained from the eastern Meditermoast of Turkey. The PCB levels in living organisms are measured from the very low, and in most cases to the detection limits. The DDE and DDT values were they high compared to PCBs and there was a linear relation between the organochlorine residue consentions and the extractable organic material of the analysis and the extractable organic material of the analysis.

meaconsiderable amount of effort has been spent in the emination of the levels of organochlorine residues in addition are water and sediments of the Mediterranean

(IAEA, 1976). However, very little work has been done in the eastern Mediterranean, especially on the Turkish coast (Balkas *et al.*, 1978).

In the present work organochlorine residues, i.e. chlorinated pesticides and polychlorinated biphenyls, present in marine organisms caught on the eastern Mediterranean coast of Turkey, as well as sediment samples obtained nearby, have been investigated. The location of the samples is shown in Fig. 1. The bony fishes selected were those which have commercial value and may be found throughout the area. The species studied were, *Mugil auratus* (golden grey mullet), *Mullus barbatus* (striped mullet), *Mullus surmuletus* (red mullet) and *Upeneus mollucensis* (gold band goat fish), *Panaeus kerathurus* (shrimp) and *Patella caerulea* (limpet) were also analysed, since they have different feeding habits to bony fishes. *Patella caerulea* is a

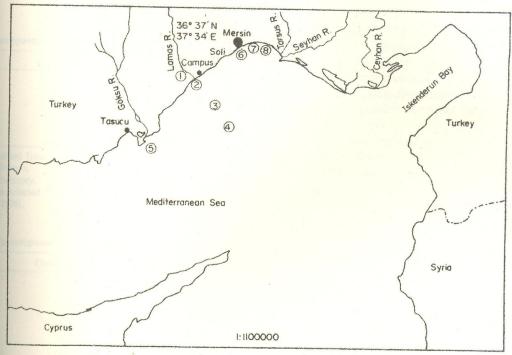


Fig. 1 Location of sampling stations on the eastern Mediterranean coast of Turkey.

Our feeder (Riedl, 1973) so it might be an indicator of the bes of organochlorine residues in sea water (Goldberg et 4,1978).

## Samples and Methods

The samples of fish and shrimp were caught by gill nets addentified according to the FAO identification Directory FAO, 1973). Limpets were collected from the rocks with the aid of metal knives. Samples were rinsed with distilled rater, weighed and measured then wrapped in ethanol-measured aluminium foil and frozen at  $-25^{\circ}$ C until adysis.

Sediment samples were taken with the aid of a grab ampler and immediately transferred into ethyl alcohol measured glass jars and finally deep frozen until analysis.

About 10 g of the muscle tissue of the living organisms ramixed with an equal amount of anhydrous Na<sub>2</sub>SO<sub>4</sub> and whet extracted with 75 ml of hexane which had been sailed over NaOH in an all-Pyrex 1 m fractional faillation column. After at least 20 cycles of extraction, of the aliquot was used for EOM (extractable organic werial) determinations. The other half of the extract was based as described elsewhere (Holden & Marsden, 1969), wolume reduced to 1 ml and eluted with 45 ml of hexane to man aluminium column. In order to separate PCBs and the sufficient, 1971) was necessary.

Thehomogenized wet sediment samples were divided into be portions. One of the portions was dried in an oven to be portions. One of the portions was dried in an oven to be portion. One of the portions was dried in an oven to be portion was extracted as haker with  $3 \times 100$  ml of methanol-benzene mixture below volume). The combined extracts were partitioned to be present the portion of the portion was used for the portion of per cent EOM and the other half cleaned and the over alumina and silica columns as described for

the organisms. Mean time blank experiments were also carried out and background correction for the residues was made.

Sediment extracts containing sulphur were treated with metallic Cu prior to GC analysis in order to remove sulphur.

The volumes of both eluates from living organisms and sediment were reduced to 1 and 2 ml, and 1  $\mu$ l aliquots were injected to GC. A Varian Model 2700 gas chromatogram equipped with a <sup>63</sup>Ni electron capture detector was used during GLC analysis. The separation was achieved by a 3 m (2 mm i.d.) coiled glass column filled with 1.5% OV17 and 2% OV 210 on Varaport 30 (80–100 mesh). The injection port, column and detector temperatures were 250, 225 and 275°C, respectively. Nitrogen flowing at 20 ml min<sup>-1</sup> was the carrier gas.

Identification of PCBs was achieved by matching the standard Aroclor (supplied by International Laboratory Inc., USA) peaks with the unknowns. The quantification was based on five major peaks of Aroclor 1254. The pesticide peaks were compared with the standards. Whenever necessary, i.e. especially in the case of intercalibration samples, where the PCB peaks in the chromatograms were dominant, dehydrochlorination with alcoholic KOH and oxidation with acidic Cr<sub>2</sub>O<sub>3</sub> were performed as confirmatory tests for DDTs.

Our laboratory is also participating in the intercalibration exercise organized by the International Laboratory of Marine Radioactivity (IAEA), Monaco. Tables 1 and 2 consists of analysis results obtained from the analysis of the Mediterranean sediment and fish homogenate supplied by IAEA during the intercalibration. The extraction, clean up and analysis procedure for the intercalibration were exactly the same as applied to the analysis of fish and sediment samples. As can be seen from Table 1 and Table 2 there is no systematic difference in the DDT and PCB concentrations reported by our laboratory and average values given by IAEA (IAEA, 1976).

TABLE 1

Morine compound concentrations in Mediterranean sediment. a

Residue	Concentrationb	ng g-1 dry weight
-BHC	1.8	0.1
BRHC	0.2	0.7
w'-DDE	1.9	$2.7 \pm 0.7$
-DDD	9.2	7 ± 4
W-DDT	16.1	$13 \pm 5.3$
PCB-1260	208.6	$180 \pm 39$

lample was obtained from the International Laboratory of Marine

Results of this laboratory

lisults of seven selected laboratories from IAEA intercalibration assist (IAEA, 1976).

TABLE 2
Organochlorine compound concentrations in fish homogenate.a

lesidue	Concentrationb	ng g <sup>-1</sup> dry weight
BHC .	7.8	16.0 ± 7.4
BHC	37.7	$41.3 \pm 17.4$
Altrin	11.1	$17.2 \pm 4.1$
J-DDE	56.4	$296.0 \pm 214.0$
W-DDD	30.7	$293.0 \pm 227.0$
W-DDT	40.0	$370.0 \pm 270.0$
KB-1254	1645.0	$3220.0 \pm 2830$

include obtained from the International Laboratory of Marine laboratory, Monaco.

lealts of this laboratory.

beil average values reported by the participating laboratories [Shoratories reporting < values have been excluded) (IAEA, 1978).

## **Rsults and Discussion**

Theresults obtained from the analysis of marine biota are min Table 3. As can be seen from this table, there was no quetable amount of PCB in any of the species analysed, and Patella caerulea which showed up to 39 ng g<sup>-1</sup> (fresh table) of PCB concentration. Table 4 consists of the results at the PCB levels were low. This observation is assent with the results reported by Elder & Villeneuve Th. They observed a decreasing trend in the PCB accuration from the western towards the eastern

Mediterranean. Harvey et al. (1974a) and Bidelman & Olney (1974) explained low dissolved PCB values in the Sargasso Sea by a co-distillation process. In the eastern Mediterranean evaporation also exceeds precipitation and co-distillation of PCBs is also possible in this area. The lack of local sources is possibly the main factor accounting for low PCB concentrations.

The t-DDT concentrations in all analysed marine biota and sediment samples are EOM dependent. With increasing EOM, the t-DDT concentration increases. When, using the 'least square method' the mean t-DDT values given in Table 3 for different species were plotted against % EOM, a straight line (Fig. 2) with a linear correlation coefficient of 0.83 and a standard error  $\pm 1.1$  was obtained. In the case of sediment samples too, with the exception of the result from Station No. 8, t-DDT vs % EOM gave a straight line (Fig. 3). The linear correlation coefficient in this case was 0.97 and the standard error  $\pm 0.25$ . The sampling point No. 8 was very close to an industrial area where agrichemical, petrochemical and some other complexes are located. This might be a reason for the relatively high t-DDT concentrations with respect to % EOM.

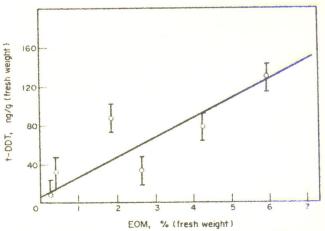


Fig. 2 t-DDT in living organisms as a function of % EOM. Curve fitting has been done by using 'least-square method'.

TABLE 3
Organochlorine concentrations in the living organisms.

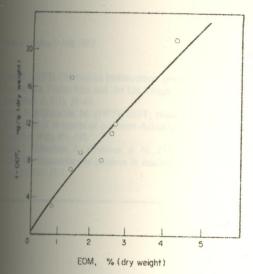
Species analysed	No. of individuals % EOM <sup>a</sup> analysed		Organochlorine residue (ng g <sup>-1</sup> F.W.)			
				t-DDE <sup>d</sup>	t-DDT	t-PCB <sup>t</sup>
Mugil auratus (golden grey mullet)	20	1.8	min max. mean	5 173 48	8 324 89	T <sup>c</sup> 10
Mullus barbatus (striped mullet)	26	6.0	min. max. mean	122 62	9 257 130	T 2
Mullus surmuletus (red mullet)	6	2.6	min. max. mean	7 35 21	20 49 34	T
Upeneus mollucensis (gold band goat fish)	30	4.3	min. max. mean	31 69 47	49 94 74	T T
Parapaneus kerathurus (shrimp)	25	0.4	min. max. mean	3 61 28	4 65 34	T T T
Patella caerulea (limpet)	42	0.3	min. max. mean	1 4 2	2 7 5	2 39 15

<sup>&</sup>lt;sup>a</sup>% extractable organic material, based on fresh weight of the living organism.

Trace indicates less than 2 ng g -1.

dt-DDE is the sum of op- and pp-DDE.

<sup>&</sup>lt;sup>b</sup>The PCB calculations were based on Aroclor 1254.



fining has been done by using 'least-square method'.

TABLE 4
Organochlorine concentrations in the sediments.

Serpting points	% EOM <sup>b</sup>	Organochlorine concentration (ng g <sup>-1</sup> dry weight)			
		1-DDEd	1-DDT	1-PCB	
	4.3	5	21	Te	
	2.3	4	8	T	
	1.7	3	9	2	
	1.4	3	7	T	
	0.8	2	3	Т	
	2.6	8	11	T	
	2.7	3	12	3	
	1.4	4	17	4	

First sampling points refer to Fig. 1.

Lead on the dry weight of the sediments. The indicates less than  $2 \text{ ng g}^{-1}$ .

Lead on the dry weight of the sediments. The indicates less than  $2 \text{ ng g}^{-1}$ .

the t-DDT concentrations observed in Mugil auratus in the eastern Mediterranean are comparable with those parted from the North Adriatic. Picer et al. (1978) and bedante & Gilmartin (1975), reported t-DDT values of the two 28 and 409 ng g<sup>-1</sup>.

Miles barbatus, Mullus surmuletus and Upeneus sacensis form the family Mullidae. Satsmadjis & and des (1977, 1979) reported t-DDT concentrations of men 9 and 390 ng g-1 in Mullus barbatus caught from anicos Bay, Greece. On the other hand, Reimold (1975) and striped mullet from the coast of Puerto Rico and kUSVirgin Islands and reported only PCBs but no DDT. MADDT concentrations from Saranicos Bay are fairly to the results given in Table 3. Harvey et al. (1974b) level lower t-DDT concentrations in shrimps from the Bank (Atlantic Ocean) than we have observed in ins from the Eastern Mediterranean; while Giam et al. Deported t-DDT concentrations about 10 times higher in the Caribbean Sea. The only available data about was reported by Robinson et al. from the Northumberland coast, who showed them contain about one third of the t-DDT found in this work. stageit is too early to say anything about the indicator inter of Patella caerulea; further research is necessary. headDT concentrations in the sediment samples given

in Table 4 are in the same range as those reported from Southern Greece (Dexter & Pavlou, 1973). In a sequence of deep sea sediments, from 10°N to 43°N in the North Atlantic, due to degradation of DDTs, only PCBs have been observed (Harvey & Steinhauer, 1976).

From the above comparisons, it is readily seen that the *t*-DDT content of living organisms and sediment samples from the eastern Mediterranean coast of Turkey were not much different from those obtained from other parts of the Mediterranean basin. However, in general the *t*-DDT concentrations are higher than those reported from the Atlantic Ocean. On the other hand, PCB concentrations are much lower than those obtained from the Mediterranean Sea and the North Atlantic Ocean.

In all samples, except the sediment sample from Station No. 8 the main constituents of the *t*-DDT was *pp'*-DDE. In the living organisms and in sediments *pp'*-DDE formed about 50% and 30% or more of the total *t*-DDT, respectively. This difference was probably due to metabolic degradation of *pp'*-DDT into *pp'*-DDE (Plimmer *et al.*, 1968).

This work was partially supported by the United Nations Environmental Programme (UNEP) as a part of the joint FAO-UNEP coordinated project of Pollution in the Mediterranean. UNEP Project No. FB 0503-75-07.

Balkas, T. I., Salihoglu, I., Tuncel, G., Tugrul, S. & Ramelow, G. (1978). Trace Metals and Organochlorine Residue Content of Mullidae Family Fishes and Sediments in the Vicinity of Erdemli (Icel), Turkey. IVes Jurnees Etud. Pollutions, pp. 159–163, Antalya, CIESM.

Bidelman, T. F. & Olney, C. E. (1974). Chlorinated hydrocarbons in the Sargasso Sea, atmosphere and surface water. Science, N. Y., 183, 516-518.

Dexter, R. N. & Pavlou, S. F. (1973). Chlorinated hydrocarbons in sediments from southern Greece. Mar. Pollut. Bull., 4, 188–190.

Elder, D. L. & Villeneuve, J. P. (1977). Polychlorinated biphenyls in the Mediterranean Sea. Mar. Pollut. Bull., 8, 19-22.

FAO of the United Nations. (1973). FAO Species Identification Sheets for Fishery Purposes, Mediterranean and Black Sea, Vol. I and Vol. II, Rome.

Giam, C. S., Hanks, A. R., Richardson, R. L., Sackett, W. M. & Wong, M. K. (1972). DDT, DDE and polychlorinated biphenyls in biota from the Gulf of Mexico and Caribbean Sea. *Pestic. Monit. J.*. 6 (3), 139–143.

Goldberg, E. D., Bowen, V. T., Farrington, J. W., Harvey, G., Martin, J. H., Parker, P. L., Risebrough, R. W., Robertson, W., Schneider, E. & Gamble, E. (1978). The Mussel Warch. Envir. Conservation, 11, 101-125.

Harvey, G. R., Steinhauer, W. G. & Miklas, H. P. (1974a). Decline of polychlorinated biphenyl concentrations in North Atlantic surface water. Nature, Lond., 252, 387-388.

Harvey, G. R., Miklas, H. P., Bowen, V. T. & Steinhauer, W. G. (1974b). Observations on the distribution of chlorinated hydrocarbons in Atlantic Ocean organisms. Mar. Res., 32, (2), 103-118.

Harvey, G. R. & Steinhauer, W. G. (1976). Biogeochemistry of PCB and DDT in the North Atlantic. *Environmental Biochemistry* (Edited by Jerome Nriagu). Ch. 15, pp. 203–221. Ann Arbot Science Publications.

Holden, A. V. & Marsden, K. (1969). Single stage clean-up of animal tissue extract for organochlorine residue analysis. J. Chromatogr. 44, 481-492.

IAEA-187 (1976). Activities of the International Laboratory of Marine Radioactivity, 1976 Report, pp. 106 (June), Monaco.

Picer, M., Picer, N. & Ahel, M. (1978). Chlorinated insecticide and PCB residues in fish and mussels of east coastal waters of the middle North Adriatic Sea, 1974–1975. Pestic. Monit. J., 12 (3), 102–112.

Plimmer, J. R., Kearney, P. C. & Von Endt, D. W. (1968). Mechanism of conversion of DDT to DDD by aerobacter aerogenes. J. Agric. Food Chem., 16, 594-597.

Reidl, R. (Ed.) (1973). Fauna and Flora der Adria, pp. 393. Verlag Paul Parey, Hamburg, Berlin. accepted, R. J. (1975). Chlorinated hydrocarbon pesticides and mercury accepted biota, Puerto Rico and the US Virgin Islands, 1972–1974. Helic Monit. J., 9 (1), 39–43.

Peac. Monit. J., 9 (1), 39-43.
Indiane, N. & Gilmartin, M. (1975). DDT, related compounds and PB in tissues of 19 species of northern Adriatic commercial fishes.

January 1972–1974.

bisson, J., Richardson, A., Crabtree, A. N., Coulson, J. C. & Potts, 6. R. (1967). Organochlorine residues in marine organisms. *Nature*, 100d, 214, 1307-1311.

Satsmadjis, J. & Gabrielides, G. P. (1977). Chlorinated hydrocarbons in striped mullet (Mullus barbatus) of Saranikos Bay. Thalassographica, 1 (2), 151-154.

Satsmadjis, J. & Gabrielides, G. P. (1979). Observations on the concentration levels of chlorinated hydrocarbons in a Mediterranean fish.

Mar. Pollut. Bull., 10, 109-111.

Snyder, D. & Reinert, R. (1971). Rapid separation of polychlorinated biphenyls from DDT and its analogues on silica gel. Bull. envir. contam. Toxicol., 6 (5), 385-390.